

Secondary organic aerosol from ozone-initiated reactions with terpene-rich household products

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Abstract

We analyzed secondary organic aerosol (SOA) data from a series of small-chamber experiments in which terpene-rich vapors from household products were combined with ozone under conditions analogous to product use indoors. Reagents were introduced into a continuously ventilated 198 L chamber at steady rates. Consistently, at the time of ozone introduction, nucleation occurred exhibiting behavior similar to atmospheric events. The initial nucleation burst and growth was followed by a period in which approximately stable particle levels were established reflecting a balance between new particle formation, condensational growth, and removal by ventilation. Airborne particles were measured with a scanning mobility particle sizer (SMPS, 10 to 400 nm) in every experiment and with an optical particle counter (OPC, 0.1 to 2.0 μm) in a subset. Parameters for a three-mode lognormal fit to the size distribution at steady state were determined for each experiment. Increasing the supply ozone level increased the steady-state mass concentration and yield of SOA from each product tested. Decreasing the air-exchange rate increased the yield. The steady-state fine-particle mass concentration ($\text{PM}_{1.1}$) ranged from 10 to $> 300 \mu\text{g m}^{-3}$ and yields ranged from 5% to 37%. Steady-state nucleation rates and SOA mass formation rates were on the order of $10 \text{ cm}^{-3} \text{ s}^{-1}$ and $10 \mu\text{g m}^{-3} \text{ min}^{-1}$, respectively.

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1. Introduction

Terpenes and ozone are commonly present indoors, and their reactions can produce particles (Weschler and Shields, 1999; Long et al., 2000; Wainman et al., 2000; Sarwar et al., 2004; Liu et al., 2004). Consumer products, such as cleaning agents and air fresheners, are common sources of terpenes (Nazaroff and Weschler, 2004). Ozone is routinely present indoors because of ventilation with ozone-containing outdoor air (Weschler, 2000).

Particle inhalation raises health concerns (Pope and Dockery, 2006). Deposition in the lungs is size-dependent (Yeh et al., 1996; Asgharian and Price, 2007), and the health effects associated with aerosol exposure depend on particle size and concentration (Oberdörster, 2001; Peters et al., 1997). The exposure impacts of indoor pollutants are amplified because (a) people spend a high proportion of their time indoors, (b) emissions that occur indoors are diluted into confined volumes and removed at slow ventilation rates, and (c) people tend to be in close proximity to indoor pollutant sources (Nazaroff, 2008). Consequently, it is important to characterize both the source strength and size distribution of significant indoor particle sources.

Secondary organic aerosol (SOA) formation from ozone-terpene interactions has been widely studied in relation to atmospheric organic aerosol formation. Those studies have uncertain direct applicability for elucidating SOA formation indoors. Several studies have measured SOA formation and growth from ozone reactions with pure terpenes or with terpene-containing products under indoor-relevant conditions. Most of these studies measured particles using an optical particle counter (Weschler and Shields, 1999; Wainman et al., 2000; Weschler and Shields, 2003; Sarwar et al., 2003; Sarwar et al., 2004; Hubbard et al., 2005; Singer et al.,

2006a). Optical particle counters (OPC) typically measure only particles that are 100 nm or larger and thus cannot characterize the ultrafine particles that are an essential component of particle nucleation and growth. A few studies have characterized ultrafine particles using a scanning mobility particle sizer (SMPS) to investigate nucleation and growth in these cases: (a) use of a pine-oil based cleaner in the presence of ozone (Long et al., 2000); (b) peeling of oranges in the presence of ozone (Vartainen et al., 2006); and (c) adding limonene to an office environment in which an ionizing air purifier was used (Alshawwa et al, 2007).

In the present study, particle formation and size-distribution dynamics are investigated with an SMPS and an OPC for experiments in a small, flow-through chamber. Ozone reacted with vapor emissions of terpene-containing consumer products (two cleaning products and an air freshener) at indoor-relevant conditions. A previous paper reported the consumption of primary constituents and the formation of secondary products from these experiments, emphasizing gaseous species (Destailats et al., 2006). Here, particle data from the same experiments are presented in detail and analyzed to characterize particle size distributions. The effects on SOA production of factors such as air-exchange rate and ozone level are investigated. Particle formation and growth dynamics are analyzed, and the particle mass formation rate is modeled.

2. Experimental conditions

Experiments were conducted in a Teflon-lined 198-L rectangular chamber. Details of the experimental set-up and gas-phase chemical analyses were presented in Destailats et al. (2006). Briefly, cleaning product vapor was continuously introduced into the chamber and, after a steady-state condition was reached, continuous ozone addition commenced. The product vapor was introduced through Teflon tubing in one bottom corner of the chamber while ozone was introduced through Teflon tubing in the diagonal opposite bottom corner. Particle sampling was

done at the middle of the chamber ceiling. Experiments were performed at 23.0 ± 0.5 °C.

Three cleaning products were tested: a pine oil cleaner (POC); an orange-oil degreaser (OOD); and a heated, scented-oil air freshener (AFR). The reactive constituents likely to contribute to SOA formation are listed here; detailed composition of the products is reported in Singer et al. (2006b). The OOD contained only one terpene, d-limonene. The POC contained several volatile and reactive constituents, including d-limonene, terpinolene, α -terpinene and α -terpineol. The AFR was the most complex mixture, with more than 30 volatile terpenes, terpenoids, and other compounds, the most reactive of which were d-limonene, linalool, dihydromyrcenol, β -citronellol, and linalyl acetate.

For each product, three experiments were conducted with two ozone supply levels and two air-exchange rates (AERs) in the following configurations: 130 ppb, 3 h^{-1} (denoted HH); 60 ppb, 3 h^{-1} (denoted MH); and 130 ppb, 1 h^{-1} (denoted HL). Seven additional experiments with the POC were executed. Conditions for each experiment are summarized in Table 1. The total cleaning product constituent levels were similar for each of the three configurations and the total amount of reactive terpenes and terpenoids in the inflow were also similar for each cleaning product, approximately 700 ppb. All experiments, unless otherwise stated, were performed with zero grade air (Airgas) humidified by means of a sparger to 50% relative humidity (RH). The zero air was virtually particle free, but flowing the air through the sparger containing pellets of activated carbon (~ 0.5 cm in diameter) introduced “seed” particles. With the sparger in use, the supply air contained a particle concentration of $\sim 400 \text{ cm}^{-3}$. The geometric median diameter was ~ 30 nm, the geometric standard deviation was 1.3, and the mass concentration was $\sim 0.005 \mu\text{g m}^{-3}$.

With the pine-oil cleaner, the range of ozone supply levels was extended (POC-VH and POC-LH) and a replicate experiment was conducted (POC-HH1 and POC-HH2). Four

supplemental experiments were also conducted. In POC-Rev, the cleaning product vapor was introduced into the chamber that already contained a steady-state level of ozone. In POC-NO_x, a steady-state level of 74 ppb of NO₂ and 1.75 ppb of NO was present in the chamber, together with the VOC mixture and before addition of ozone, to explore the effect of the nitrate radical on oxidative chemistry. In POC-Seed, laboratory air was used instead of zero air to investigate the effects of a more realistic atmospheric seed particle distribution on secondary product formation. The POC-NO_x, POC-Rev, POC-Seed experiments were performed at an AER of 3 h⁻¹ and ~130 ppb supply ozone. To investigate the effect of water vapor, a POC-Dry experiment was performed using zero-air without any humidification at 3 h⁻¹ and ~60 ppb of ozone in supply air.

Tracer gas tests confirm that the chamber was well mixed (Destailats et al., 2006). A comparison of estimated characteristic times for mixing and reaction indicates that these two processes have similar time scales. Particles were only sampled from one position in the chamber. Some features of the particle data could be influenced by spatial variability given the burst nature of nucleation.

3. Particle measurement and analysis

3.1. Instruments

Aerosol measurements were performed using a scanning mobility particle sizer (SMPS) in every experiment and an optical particle counter (OPC) in some experiments. The SMPS measured particles in the diameter range 0.008-0.415 μm in 64 bins; data from 0.01 to 0.4 μm were used for analysis in this paper. The SMPS consists of a differential mobility analyzer (3701A, TSI Inc.) and a condensation particle counter (3760, TSI Inc.). The SMPS performed a complete scan (up and down the size distribution) approximately every two minutes. The data were collected and inverted using the Labview interface with software written by D Collins

(Texas A&M University) and P Chuang (UCSC) and analyzed using Igor (Wavemetrics Inc.) with custom routines. The OPC (Lasair 1003, Particle Measuring Systems, Inc.) measured 0.1-2 μm diameter particles in 8 bins.

For both instruments, particle volume concentration was estimated by multiplying the measured number concentration by $(\pi/6 \times \text{GMD}^3)$, where the geometric mean diameter of a bin, GMD, is the square root of the product of the upper and lower bin diameters. A particle density of 1 g cm^{-3} was assumed in converting volume to mass; this may underestimate particle mass concentration as some studies have reported the density of organic atmospheric particles to be 1.2 to 1.5 g cm^{-3} (Khlystov et al., 2004; Turpin and Lim, 2001).

The OPC sampled for 1 minute every 2 minutes, counting particles in eight bins simultaneously. The lower bin bounds, as calibrated by the manufacturer with polystyrene latex (PSL) particles, were 0.1, 0.2, 0.3, 0.4, 0.5, 0.7, 1.0, and $2.0 \mu\text{m}$. However, the instrument's response is influenced by the particle's refractive index, m , which depends on its chemical composition. The refractive index of PSL is 1.588. Organic particles tend to have a lower refractive index (Dick et al., 2007). Accurately sizing particles with a different refractive index requires scaling the bin bounds. For instance, a $0.15 \mu\text{m}$ particle of oleic acid ($m=1.46$) would be sized as a $0.1 \mu\text{m}$ particle in an OPC calibrated using PSL (Hand and Kreidenweis, 2002).

3.2. Alignment routine

Hand and Kreidenweis (2002) calibrated the same model of OPC as used in this study with PSL ($m=1.588$), dry ammonium sulfate ($m=1.53$), and oleic acid ($m=1.46$). From those data, they developed polynomials for scaling the manufacturer's bin bounds. In the present study, a routine was developed that employed these scaling polynomials in aligning the data collected where the size range measured by the OPC and SMPS overlap. At each time step, the alignment

routine scanned the range of possible m values from 1.46 to 1.59, using increments $\Delta m = 0.01$, calculated OPC bin bounds, summed the SMPS number concentrations within those bounds, and compared the results to the OPC number concentration. The result was a time-dependent m value that produced optimal alignment between the OPC and SMPS data, determined using the least-squares difference between the SMPS and adjusted OPC concentrations. Only two or three size bins of the OPC data overlapped with the size range of the SMPS data. The refractive index determined from this routine is not intended to be a robust indication of the refractive index of the particles. Instead, the purpose was to determine an appropriate adjustment of the OPC bin bounds, to improve estimates of particle mass concentrations.

The results matched our expectation that SOA has a refractive index similar to that of oleic acid. The alignment routine indicated $m = 1.46$ - 1.49 . The OPC bins were scaled for oleic acid and the lower bin bounds used were 0.15, 0.24, 0.36, 0.47, 0.62, 0.89, and 1.1 μm . (The eighth bin was not modeled.) For these seven bins, total mass concentration is equivalent to $\text{PM}_{1.1}$ because the maximum particle size detected was less than 1.1 μm .

3.3. Wall losses

The pseudo first-order rate coefficient for particle deposition to chamber surfaces, L_{dep} , was determined using equation 1 at each particle scan.

$$L_{dep} = \int_{D_{p,\min}}^{D_{p,\max}} (\beta_{dep}(D_p)) \left(\frac{dN}{d \log D_p} \right) d \log D_p \quad (1)$$

The size-dependent deposition loss-rate coefficients, β_{dep} , were estimated using the model of Lai and Nazaroff (2000). This model requires the input of a friction velocity to characterize near-surface flows. The actual friction velocity in the chamber was unknown, so a range of plausible values was used. The calculated characteristic time for deposition was very long compared to

removal by ventilation. For example, in experiment POC-MH, the minimum characteristic time for surface deposition throughout the duration of the experiment (which occurs at the peak number concentration) was 70 to 700 h, respectively, for friction velocities of 3 to 0.3 cm s⁻¹. The characteristic time for particle loss by ventilation was 0.3 h. Consequently, we concluded that particle deposition to chamber surfaces could be neglected in further analysis.

Characteristic times for deposition of condensable vapors to walls were also calculated using the Lai and Nazaroff (2000) model, but using a range of diffusion coefficients for semi-volatile products of terpene oxidation (0.04 to 0.08 cm² s⁻¹). With a minimal diffusion coefficient of 0.04 cm² s⁻¹ and the same range friction velocities used to calculate particle deposition (3 to 0.3 cm s⁻¹), the characteristic time for loss of vapors to walls was 200 to 2000 s. Although this is fast compared to ventilation it is a relatively small sink compared to condensation on particles. The first-order condensation rate for vapor onto particles, L_{cond} , was estimated using the Fuchs and Sutugin approach to mass transfer of a gas to a particle in the transition regime. The condensation coefficient, β_{cond} , was calculated using equation 2 (Equation 11.34 in Seinfeld and Pandis, 1998) and integrated over the particle size range as shown in equation 1.

$$\beta_{cond}(D_p) = 2\pi D_p D_g \left(\frac{1 + Kn}{1 + 1.71Kn + 1.33Kn^2} \right) \quad (2)$$

The diffusion coefficient is D_g , and Kn is the Knudsen number. As an example, in experiment POC-MH, the characteristic time for vapor deposition to particles, assuming a diffusion coefficient of 0.04 cm² s⁻¹ was 6 s just after the initial nucleation event and a maximum of 10 s for the remainder of the experiment. For the range of conditions and vapor diffusivities anticipated in these experiments the time scale for condensation to particles was at least an order of magnitude lower than to walls.

3.4. Coagulation sink

The pseudo first-order rate coefficient for particle loss by coagulation, L_{coag} , was determined using equation 3 for each scanned particle size distribution.

$$L_{coag}^* = \int_{D_{p,min}}^{D_{p,max}} \beta_{coag}(D_p) \left(\frac{dN}{d \log D_p} \right) d \log D_p \quad (3a)$$

$$L_{coag} = \int_{D_{p,min}}^{D_{p,max}} L_{coag}^*(D_p) \left(\frac{dN}{d \log D_p} \right) d \log D_p \quad (3b)$$

The size-dependent loss coefficients, β_{coag} , for coagulation were calculated using the Fuchs form of the Brownian coagulation coefficient as given in Table 12.1 of Seinfeld and Pandis (1998). Compared with removal by ventilation, coagulation was estimated to contribute ~25% to total particle number concentration sink immediately after nucleation and ~10% during later stages.

4. SOA formation and growth stages

SMPS data from several characteristic POC experiments are presented in Figure 1. Similar particle formation and growth behavior was exhibited in all experiments, and this behavior is divided into four stages for the present discussion. The stages are illustrated in Figure 2 using experiment AFR-HH as an example. Stage 1 is characterized by an initial nucleation burst occurring immediately after ozone is introduced. The dominant feature of stage 2 is the growth of the particles formed in the initial burst. Stage 3 occurs after ventilation has removed the majority of the particles created in the original nucleation burst; this stage is characterized by the onset of secondary nucleation events. In stage 4, all of the original particles have left the system and the system has reached a quasi-steady-state. This stage is referred to as quasi-steady-state because the particle size distribution exhibits persistent time-dependent behavior, but of a repetitive or cyclic character. Particle size distribution trends are discussed in §5 and formation

and growth dynamics are addressed in §6.

5. Size distribution characteristics

Figure 3 shows examples of the particle size distributions from stage 4 of three experiments. Figure 3a is an example where both the SMPS and OPC data could be combined to produce a good fit. Figure 3b shows an experiment where almost the entire distribution was within the range of the SMPS. Figure 3c shows an example of an experiment where the distribution was outside the range of the SMPS but no OPC data were available. In these cases, the characteristics of the upper size range of the distribution were estimated using typical distribution parameters determined from other experiments.

In experiments where both SMPS and OPC data were collected, lognormal distributions were fit to the combined (SMPS and adjusted OPC) size distributions. Three modes were required to fit the data well; using four modes did not significantly improve the fit. Manual fitting of the lognormal distribution parameters was required because, mathematically, the measured distributions were not well constrained owing to missing data at the tails. Modeled distribution parameters (number concentration, N ; geometric mean diameter, GMD; and geometric standard deviation, GSD) are presented in Table 2.

For experiments without OPC data, the height and central tendency (N and GMD) for the largest diameter mode could be determined with reasonable confidence from the SMPS data, but the spread (GSD) of the distribution was unknown. Since most fits with both OPC and SMPS data indicated a GSD of 1.4 for the third mode, this value was used to fit the rest of the experiments where only SMPS data were available. Overlaid in each example in Figure 3 is the best-fit 3-mode lognormal distribution.

Table 2 presents the measured and modeled particle data for each of the small chamber

experiments. Peak total number concentrations during stage 2 were on the order of 10^5 cm^{-3} and the particle number concentrations during stage 4 were an order of magnitude lower. Since the vast majority of the particles were smaller than 400 nm, the SMPS provided a fairly accurate measure of the total particle number concentration. Mass concentration ($PM_{1.1}$) ranged from tens to low hundreds of $\mu\text{g m}^{-3}$, and the stage 4 mass concentration was about half of the value at the peak. Together, the SMPS and OPC captured the full range of particle sizes, but up to half of the mass went undetected when only the SMPS was used.

5.1. Effect of ozone level and air-exchange rate on mass concentration and yield

Aerosol yield (Y) was calculated as the total mass concentration of SOA ($PM_{1.1,SS}$) formed per mass of VOC consumed (ΔVOC). Integrated gas-phase samples were taken at steady-state conditions before and after ozone addition, so yields were calculated using steady-state particle mass concentration. For the two experiments (OOD-HL and POC-ML) where the size distribution varied significantly during the steady-state period, the average $PM_{1.1,SS}$ was used to estimate yield.

As shown in Figure 4a, the steady-state mass concentration of SOA was greater at the higher supply ozone level for each of the three products. The relationship between yield and ozone level was similar to the relationship between mass concentration and ozone, although the yield for POC appeared to level off above ~60 ppb ozone, as shown in Figure 4b. Limonene, α -terpinene, and terpinolene have the highest SOA-forming potential of the terpenes in the tested products (Lee et al., 2006; Koch et al., 2000). The total terpene level (in ppb) introduced was roughly constant from one experiment to the next, but the fraction of terpenes with high SOA-forming potential decreased from 100% in OOD to 65% in POC to 10% in AFR. The relative level of SOA production for each household product roughly corresponds to these ratios.

The yield values agree with other indoor studies. A yield of 10-15% was estimated from introducing a limonene source into an office ventilated with outdoor air that contained a moderate amount of ozone (Weschler and Shields, 1999). A yield of ~25% was calculated from particle and terpene data collected from a pine-scented heated air freshener in a chamber study with 50 ppb residual ozone (Liu et al., 2004, yields calculated from Figure 7). A yield of 13% was reported for limonene injected into an office in which an air-cleaning device that produces ozone as a byproduct was operated (Alshawwa et al., 2007).

Relative to the case for $AER = 3 \text{ h}^{-1}$, SOA production appears to decrease for POC and OOD and increase for AFR when $AER = 1 \text{ h}^{-1}$ (see Figure 4a). However, this plot obscures the relationship between particle production and air-exchange rate. When the steady-state particle concentration is compared to the mass rate of precursor consumption (Figure 4c), it is clear that more SOA is produced per mass of precursor consumed when the air-exchange rate is lowered. This finding is consistent with the expectation that longer reaction times allow for greater oxidation of the ozone reaction products. Several studies have shown that second-generation products of the ozone-limonene reaction make significant contributions to total SOA (Leungsakul et al., 2005; Ng et al., 2006; Zhang et al., 2006). The increased time in the chamber may also allow for more oxidation of the slower-reacting compounds, especially in AFR.

5.2. Effect of RH, order of reagent addition, and ambient seed particles

The POC-Rev, POC-NO_x, and POC-Seed experiments were conducted under the same conditions as POC-HH1 and POC-HH2. Experiment POC-Rev exhibited, as expected, little change in the stage 4 particle characteristics. It also had little effect on the nucleation burst and initial growth, which can be seen by comparing Figures 1a and 1d. The addition of NO_x to the supply air did not have an evident effect on SOA formation. In POC-Seed, the seed particle

number concentration was similar to that in humidified zero-air experiments ($\sim 500 \text{ cm}^{-3}$), but the distribution of seed particles was shifted toward larger particles (GMD = 98 nm, GSD = 2.2) compared with zero air (GMD = 30 nm, GSD = 1.3), resulting in 100 times greater particle mass concentration in the supply air (0.5 versus $0.005 \text{ } \mu\text{g m}^{-3}$). The steady-state particle number and mass concentrations were $\sim 2\text{-}3\times$ and $\sim 35\%$ higher, respectively, for POC-Seed than POC-HH1.

The POC-Dry experiment was conducted under the same conditions as POC-MH. In running the “dry” experiment, we discovered that the water sparger used to humidify air was the source of seed particles in all 50% RH experiments. Cocker et al. (2001) reported that moderate RH tended to increase the overall aerosol yield owing to the hygroscopicity of aerosol-phase organic material, whereas aqueous seed particles containing salts would tend to lower the overall aerosol yield owing to interactions between the salts and the organic material. While the seed particles in our experiments were likely aqueous, it is unknown whether they contained salts.

The effects of removing the seed particles and lowering the RH cannot be separated in our experiments; the overall effect was to lower the particle number and mass concentration (see Figures 1b and 1c). The number and mass particle concentrations in the supply air for the POC-Dry experiment were $< 5 \text{ cm}^{-3}$ and $< 0.001 \text{ } \mu\text{g m}^{-3}$, respectively, and corresponding concentrations in POC-MH were 380 cm^{-3} (GMD = 24 nm, GSD = 1.3) and $0.005 \text{ } \mu\text{g m}^{-3}$.

6. Particle formation and growth characteristics

6.1. Cycle of particle formation and growth

Figure 1f shows experimental results where a repeated cycle of particle formation and growth occurs in stages 3 and 4. The balance between new particle formation and growth of existing particles in the chamber can be plausibly explained as follows. As particles grow by condensation, they are also removed by ventilation, reducing the total surface area available

for condensation. Condensable vapor then accumulates until a burst of nucleation occurs that significantly lowers the vapor concentration. These newly formed particles grow while continuing to be removed from the chamber by ventilation, again reducing the surface area available for condensation, and the cycle repeats. The balance between nucleation and growth is most clearly visible in the low AER experiments, as illustrated by the repeated appearance of plumes in Figure 1f for experiment POC-HL. In contrast, in the higher AER experiments, particles are so rapidly ventilated out of the system that new particle formation by nucleation seems to occur almost constantly. The plumes of these nucleation events are compressed to a timescale similar to that of the SMPS measurements, giving the appearance of a steady-state particle size distribution. An example of this behavior is illustrated in Figure 1a for experiment POC-HH. Experiment POC-MH exhibited behavior between these two extremes, as shown in Figure 1b. Trends in particle mass and number concentration support this description. In experiments where the particle behavior is dynamic, such as in POC-HL (Figure 1f) and OOD-HL, the mass concentration and GMD increase and the number concentration decreases after each nucleation event.

This process of condensable vapor accumulation and subsequent nucleation burst is also thought to be responsible for the “fingers,” or separate peaks in the size distribution that appear in stage 2 of every experiment. Once the initial burst of particles is formed, some are removed by ventilation or there is otherwise a build up of condensable vapor that cannot be accommodated by condensation on the existing particles and a small new burst occurs, creating a thin plume. Peaks grow by condensation at rates similar to those around it, and it is not until enough particle surface area is removed by ventilation that this phenomenon is halted.

The rate of decay of particle number concentration in the chamber following the initial

burst provides additional evidence of persistent particle nucleation. After the initial nucleation burst, the decay rate of the particle number concentration is less than the loss rate expected from ventilation alone. Figure 5 compares the observed pseudo first-order loss rate of particle number concentration with the air-exchange rate for experiment POC-MH. The apparent loss rate was 2.5 h^{-1} while the AER was 3.0 h^{-1} . We infer that nucleation must be occurring to provide a fresh source of new particles that offsets some of the removal by ventilation. In this experiment, about 40 minutes after ozone was introduced (at 16:25), the particle loss rate slowed, rebounded slightly and then settled at an effectively constant level. The inflection corresponds to the onset of stage 3, with the occurrence of distinct new nucleation events, as can be seen in Figure 1b. New particle formation occurred because not enough of the original particles remained in the chamber to accommodate the condensable vapor being formed. The dip in the particle number concentration is the result of two effects. First, the originally created particles have grown so large that they are no longer counted by the SMPS. Second, so long as the large particles from the initial burst remain present, their growth in surface area means that fewer particles are required to accommodate the flow of condensable mass as time progresses.

The balance between condensation and nucleation is also evident in the OPC data. As shown in Figure 2, the number concentration of particles in the $0.62\text{-}0.89 \text{ }\mu\text{m}$ size bin stagnates as secondary nucleation starts at the onset of stage 3. The stagnation occurs because the condensational growth is now apportioned between the numerous newly formed particles and the small number of residual large particles. Temporary stagnation in this size bin at the onset of secondary nucleation is seen in all experiments for which OPC data were collected, and the effect was more pronounced for experiments with higher particle concentrations.

6.2. Nucleation subsequent to the initial event

The nucleation events subsequent to the initial burst do not produce the large number concentrations of particles that were created in the initial event. There are two likely reasons for this depressed nucleation intensity. First, there is not as much condensable material in the system in stages two through four as there was when ozone was first introduced. When ozone first enters the chamber, the concentration of terpenes in the system is at its highest; for the remainder of each experiment, ozone only can react with the residual terpene concentrations. Second, there are many preexisting particles in the system after the initial burst, and some of the condensable material contributes to particle growth rather than to nucleation.

The inference that nucleation is occurring in stages 3 and 4, rather than just growth of seed particles, is reinforced by two observations: the particles present in the third and fourth stages are smaller than those present in the supply air; and the number concentration of particles in the chamber persists at a level that is orders of magnitude higher than the number concentration of particles in the supply air. It seems very unlikely that so many seed particles smaller than the lower limit of the SMPS (~8 nm) would exist in the supply air and could therefore be responsible for the appearance of smaller particles in the system through growth.

At steady state, the particle nucleation rate can be evaluated from a number balance. The only significant source contributing to the total particle number concentration in the chamber, P_N , is nucleation, and the dominant sink is removal by ventilation, so a material balance is:

$$\frac{dP_N}{dt} = R_N - \lambda P_N \quad R_N = \lambda P_{N,SS} \quad (4)$$

At steady state, the rate of nucleation, R_N , can be estimated as the chamber air-exchange rate, λ , times the steady-state particle number concentration, $P_{N,SS}$. The number of particles in the supply air is negligible compared to the number of particles when ozone and terpenes are present, and thus air supply is not included as a source. For example, in experiment POC-MH, the supply air

contained a particle concentration of 350 cm^{-3} , whereas the stage 4 chamber concentration was $24,000 \text{ cm}^{-3}$. In this experiment, the nucleation rate was estimated to be $20 \text{ cm}^{-3} \text{ s}^{-1}$. Multiplying by the volume of the chamber (198 L) gives a total particle generation rate of $4 \times 10^6 \text{ s}^{-1}$. For comparison, in a study in a lab room where limonene was released as a result of peeling oranges in the presence of ~ 20 ppb ozone, the calculated particle nucleation rate throughout the room was considerably smaller, $\sim 10^5 \text{ s}^{-1}$ (Vartiainen et al., 2006).

Similarly, a balance on particle mass concentration, P_M , considers formation from condensation of reaction products, balanced by removal by means of ventilation:

$$\frac{dP_M}{dt} = R_C - \lambda P_M \quad R_C = \lambda P_{M,SS} \quad (5)$$

Again, at steady state, the production rate of condensed material, R_C , must approximately equal the chamber air-exchange rate, λ , times the total measured steady-state particle mass concentration, $P_{M,SS}$. The mass concentration of particles in the supply air is negligible compared with the concentration when ozone and terpenes are present, and thus is not included as a source. For example, in experiment POC-MH, the supply air contained $0.005 \mu\text{g m}^{-3}$ and the stage 4 concentration in the chamber was $76 \mu\text{g m}^{-3}$. R_C represents only a fraction of the total production rate of reaction products, since some of the reaction products may persist in the gas phase or deposit on chamber surfaces. For the POC-MH experiment, the mass production rate of the particle-phase condensed material was $3.8 \mu\text{g m}^{-3} \text{ min}^{-1}$. The steady-state particle nucleation and mass production rates calculated in this way for all experiments using the steady-state number and mass concentrations from Table 2 and the AER from Table 1 were in the range $1\text{-}23 \text{ cm}^{-3} \text{ s}^{-1}$ and $1\text{-}13 \mu\text{g m}^{-3} \text{ min}^{-1}$, respectively, as reported in Table 2.

During stage 4 in many experiments, the mean particle size tended to increase slightly over several hours (see Figure 1d). In smog-chamber experiments where ozone is reacted with

terpenes, as condensable matter is formed, the GMD of the particles increases; however, in our flow-through system, where particles are continuously being created and swept out of the chamber, the explanation for growth of the GMD of the aerosol distribution is not apparent.

6.3. Modeling total particle mass formation

Can a simple mechanistic model be used to describe the rate of increase in particle mass concentration during the initial parts of these experiments? The OOD experiments were used for this exercise because OOD only contains one reactive compound, d-limonene. We used a numerical approximation to solve coupled differential equations for the species considered: ozone [O_3], limonene [L], and secondary particle mass [SOA] (equations 6-8).

$$\frac{d[O_3]}{dt} = \lambda[O_3]_{\text{supply}} - \lambda[O_3] - k_{O_3}[O_3][L] \quad (6)$$

$$\frac{d[L]}{dt} = \lambda[L]_{\text{supply}} - \lambda[L] - k_{O_3}[O_3][L] \quad (7)$$

$$\frac{d[SOA]}{dt} = \lambda[SOA]_{\text{supply}} - \lambda[SOA] + Yk_{O_3}[O_3][L] \quad (8)$$

Units of ppb were used for ozone and limonene concentrations in equations 6 and 7, but the limonene concentration was converted to units of $\mu\text{g m}^{-3}$ for use in equation 8. The SOA concentration was also expressed in units of $\mu\text{g m}^{-3}$. The reaction rate for ozone and limonene, k_{O_3} , was set to the experimentally determined value of $5.16 \times 10^{-6} \text{ ppb}^{-1} \text{ s}^{-1}$ (Hakola et al., 1994). The yield, Y , was set to the value determined in the present study (values shown in Figure 4b).

In the two cases in which the AER was relatively high (OOD-HH and OOD-MH), this model captured the overall shape of the increase in total particle mass concentration after the initial nucleation burst, but slightly underestimated particle mass. However, in the low AER case (POC-HL), the model substantially underpredicted both the total mass of SOA formed and the rate of formation (Figure 6). While the yield value taken from the experiment is an overall value

that incorporates all factors contributing to SOA formation, the model may underpredict mass formation because it considers only reactions of ozone and limonene. Other factors that may contribute to SOA mass are uptake of water by the hygroscopic SOA (Cocker et al., 2001) and condensation of “second-generation” oxidation products.

Ozone reaction with limonene produces first-generation oxidation products and the hydroxyl radical (Aschmann et al., 2002). Ozone and OH react with these oxidation products to form second-generation products and so on. Subsequent oxidation products can be even less volatile than first-generation products leading to increased SOA mass (Leungsakul et al., 2005; Ng et al., 2006; Zhang et al., 2006). We postulate that in the low AER case, the secondary oxidation products formed from the ozone-limonene reaction have additional time to react, forming even lower volatility products, which tend to condense and increase the total mass of SOA formed. This inference is consistent with the lower ozone and OH radical concentrations previously reported for the low AER experiments, as compared with those determined for the same conditions, but with higher AER (Destailats et al., 2006).

7. Conclusions

In this study, a series of chamber experiments was used to characterize the dynamics of particle formation and growth from ozone reactions with terpene-containing vapors from consumer products under indoor-relevant conditions. The particles formed were in the ultrafine and accumulation modes ($< 1.1 \mu\text{m}$). Particles were measured with an SMPS (10 nm to 400 nm) and an OPC (0.1 to $1.1 \mu\text{m}$), and data from the two instruments were aligned by means of adjusting the OPC bin bounds by fits informed by an assumed composition of the aerosol. The mass concentrations of ultrafine and accumulation mode particles ($\text{PM}_{1.1}$) formed by ozone reaction with terpene-containing cleaning products ranged from the tens to low-hundreds of μg

m^{-3} . For comparison, yearly maximum 24-hour ambient $\text{PM}_{2.5}$ concentrations are in the same range (USEPA, 2008). Hence, relative to health-based standards, this evidence suggests that indoor reactions between ozone and terpenes can be a significant source of particles, warranting further study of potential exposure-related health effects.

In each of the 16 experiments conducted, a burst of particle formation by nucleation occurred immediately after ozone addition. This burst was followed by a period characterized predominantly by condensational growth of the nucleated particles. The system then evolved through a third stage to a fourth during which particle nucleation and growth persisted under steady state or cyclic conditions for the remainder of the experiment. At higher air-exchange rates, the particle surface area was reduced rapidly as particles were swept out of the chamber, and nucleation seemed to occur constantly, whereas in the lower AER experiments an oscillating dynamic balance between formation and growth was exhibited. Mass and yield of SOA were observed to increase with increasing ozone level for the range of ozone levels likely to be encountered under normal indoor conditions. More SOA was formed per unit precursor consumed when the air-exchange rate was lowered. The additional SOA formation may be a result second-generation oxidation.

In any real environment there would likely be much greater spatial heterogeneity than in these small-chamber experiments. Nevertheless, SOA formation and growth has been measured in realistic settings and has exhibited similar characteristics (size distribution and growth dynamics) to the SOA production measured in these more controlled experiments. Although information about particle formation and growth can be ascertained from instruments with lower size and time resolution, certainly instruments with higher resolution and lower limits of particle size detection provide a clearer picture of the particle dynamic processes. Studies such as the one

presented here help to elucidate the distribution and evolution of particles that people are likely exposed to when vapors from terpene-containing cleaning products or air fresheners are simultaneously present with ozone indoors.

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References

- Alshawwa, A., Russell, A.R., Nizkorodov, S.A., 2007. Kinetic analysis of competition between aerosol particle removal and generation by ionization air purifiers. *Environmental Science & Technology* 41, 2498-2504.
- Aschmann, S.M., Arey, J., Atkinson, R., 2002. OH radical formation from the gas-phase reactions of O₃ with a series of terpenes. *Atmospheric Environment* 36, 4347-4355.
- Asgharian, B., Price, O.T., 2007. Deposition of ultrafine (NANO) particles in the human lung. *Inhalation Toxicology* 19, 1045-1054.
- Cocker III, D.R., Clegg, S.L., Flagan, R.C., Seinfeld, J.H., 2001. The effect of water on gas-particle partitioning of secondary organic aerosol. Part I: α -pinene/ozone system. *Atmospheric Environment* 35, 6049-6072.
- Destailats, H., Lunden, M.M., Singer, B.C., Coleman, B.K., Hodgson, A.T., Weschler, C.J., Nazaroff, W.W., 2006. Indoor secondary pollutants from household product emissions in the

- presence of ozone: a bench-scale chamber study. *Environmental Science & Technology* 40, 4421-4428.
- Dick, W.D., Ziemann, P.J., McMurry, P.H., 2007. Multiangle light-scattering measurements of refractive index of submicron atmospheric particles. *Aerosol Science and Technology* 41, 549-569.
- Hakola, H., Arey, J., Aschmann, S.M., Atkinson, R., 1994. Product formation from the gas-phase reactions of OH radicals and O₃ with a series of monoterpenes. *Journal of Atmospheric Chemistry* 18: 75-102.
- Hand, J.L., Kreidenweis, S.M., 2002. A new method for retrieving particle refractive index and effective density from aerosol size distribution data. *Aerosol Science and Technology* 36, 1012-1026.
- Hubbard, H.F., Coleman, B.K., Sarwar, G., Corsi, R.L., 2005. Effects of an ozone-generating air purifier on indoor secondary particles in three residential dwellings. *Indoor Air* 15, 432-444.
- Khlystov, A., Stanier, C., Pandis, S.N., 2004. An algorithm for combining electrical mobility and aerodynamic size distributions data when measuring ambient aerosol. *Aerosol Science and Technology* 38 (Suppl. 1), 229-238.
- Koch, S., Winterhalter, R., Uherek, E., Koloff, A., Neeb, P., Moortgat, G.K., 2000. Formation of new particles in the gas-phase ozonolysis of monoterpenes. *Atmospheric Environment* 34, 4031-4042.
- Lai, A.C.K., Nazaroff, W.W., 2000. Modeling indoor particle deposition from turbulent flow onto smooth surfaces. *Journal of Aerosol Science* 31, 463-476.
- Lee, A., Goldstein, A.H., Keywood, M.D., Gao, S., Varutbangkul, V., Bahreini, R., Ng, N.L., Flagan, R.C., Seinfeld, J.H., 2006. Gas-phase products and secondary aerosol yields from the ozonolysis of ten different terpenes. *Journal of Geophysical Research-Atmospheres* 111, D07302.
- Leungsakul, S., Jaoui, M., Kamens, R.M., 2005. Kinetic mechanism for predicting secondary organic aerosol formation from the reaction of d-limonene with ozone. *Environmental Science & Technology* 39, 9583-9594.

- Liu, X., Mason, M., Krebs, K., Sparks, L., 2004. Full-scale chamber investigation and simulation of air freshener emissions in the presence of ozone. *Environmental Science & Technology* 38, 2802-2812.
- Long, C.M., Suh, H.H., Koutrakis, P., 2000. Characterization of indoor particle sources using continuous mass and size monitors. *Journal of the Air & Waste Management Association* 50, 1236-1250.
- Nazaroff, W.W., Weschler, C.J., 2004. Cleaning products and air fresheners: exposure to primary and secondary air pollutants. *Atmospheric Environment* 38, 2841-2865.
- Nazaroff, W.W., 2008. Inhalation intake fraction of pollutants from episodic indoor emissions. *Building and Environment* 43, 269-277.
- Ng, N.L., Kroll, J.H., Keywood, M.D., Bahreini, R., Varutbangkul, V., Flagan, R.C., Seinfeld, J.H., Lee, A., Goldstein, A.H., 2006. Contribution of first- versus second-generation products to secondary organic aerosols formed in the oxidation of biogenic hydrocarbons. *Environmental Science & Technology* 40, 2283-2297.
- Oberdörster, G., 2001. Pulmonary effects of inhaled ultrafine particles. *International Archives of Occupational and Environmental Health* 74, 1-8.
- Peters, A., Wichmann, H.E., Tuch, T., Heinrich, J., Heyder, J., 1997. Respiratory effects are associated with the number of ultrafine particles. *American Journal of Respiratory and Critical Care Medicine* 155, 1376-1383.
- Pope, C.A., Dockery, D.W., 2006. Health effects of fine particulate air pollution: lines that connect. *Journal of the Air & Waste Management Association* 56, 709-742.
- Sarwar, G., Corsi, R., Allen, D., Weschler, C., 2003. The significance of secondary organic aerosol formation and growth in buildings: experimental and computational evidence. *Atmospheric Environment* 37, 1365-1381.
- Sarwar, G., Olson, D.A., Corsi, R.L., Weschler, C.J., 2004. Indoor fine particles: the role of terpene emissions from consumer products. *Journal of the Air & Waste Management Association* 54, 367-377.
- Seinfeld, J.H., Pandis, S.N., 1998. *Atmospheric chemistry and physics: from air pollution to*

- climate change. Wiley, New York.
- Singer, B.C., Coleman, B.K., Destailats, H., Hodgson, A.T., Lunden, M.M., Weschler, C.J., Nazaroff, W.W., 2006a. Indoor secondary pollutants from cleaning product and air freshener use in the presence of ozone. *Atmospheric Environment* 40, 6696-6710.
- Singer, B.C., Destailats, H., Hodgson, A.T., Nazaroff, W.W., 2006b. Cleaning products and air fresheners: emissions and resulting concentrations of glycol ethers and terpenoids. *Indoor Air* 16, 179-191.
- Turpin, B.J., Lim, H.J., 2001. Species contributions to PM_{2.5} mass concentrations: revisiting common assumptions for estimating organic mass. *Aerosol Science & Technology* 35, 602-610.
- USEPA (United States Environmental Protection Agency), 2008. Criteria air pollutants monitor values report for PM_{2.5} in United States in 2007.
(<http://www.epa.gov/air/data/help/hmonvals.html?us~USA~United%20States>)
- Vartiainen, E., Kulmala, M., Ruuskanen, T.M., Taipale, R., Rinne, J., Vehkamäki, H., 2006. Formation and growth of indoor air aerosol particles as a result of D-limonene oxidation. *Atmospheric Environment* 40, 7882-7892.
- Wainman, T., Zhang, J.F., Weschler, C.J., Liou, P.J., 2000. Ozone and limonene in indoor air: a source of submicron particle exposure. *Environmental Health Perspectives* 108, 1139-1145.
- Weschler, C.J., 2000. Ozone in indoor environments: concentration and chemistry. *Indoor Air* 10, 269-288.
- Weschler, C.J., Shields, H.C., 1999. Indoor ozone/terpene reactions as a source of indoor particles. *Atmospheric Environment* 33, 2301-2312.
- Weschler, C.J., Shields, H.C., 2003. Experiments probing the influence of air-exchange rates on secondary organic aerosols derived from indoor chemistry. *Atmospheric Environment* 37, 5621-5631.
- Yeh, H.C., Cuddihy, R.G., Phalen, R.F., Chang, I.Y., 1996. Comparisons of calculated respiratory tract deposition of particles based on the proposed NCRP model and the new ICRP66 model. *Aerosol Science and Technology* 25, 134-140.

Zhang, J., Huff Hartz, K.E., Pandis, S.N., Donahue, N.M., 2006. Secondary organic aerosol formation from limonene ozonolysis: homogeneous and heterogeneous influences as a function of NO_x . *Journal of Physical Chemistry A* 110, 11053-11063.

Figure Captions

- Figure 1.** Particle size-distribution evolution as measured with an SMPS for six POC experiments. The y-axis indicates particle diameter, D_p (nm), the x-axis represents time (with tick marks indicating 30-min increments) and the shading indicates the count-based particle size distribution, $dN/d(\log D_p)$ (cm^{-3}).
- Figure 2.** Example of the characteristic stages of particle formation and growth in chamber experiments, illustrated for experiment AFR-HH. Stage 1 is characterized by a sudden nucleation burst when ozone is added to a steady-state level of product vapor. Stage 2 is characterized by the growth of particles from the initial nucleation burst. In stage 3 nucleation resumes but particles from the initial nucleation event are still present. In stage 4 a steady or oscillating particle concentration is achieved. The color scale for $dN/d(\log D_p)$ in the lower frame is the same as in Figure 1.
- Figure 3.** Size distributions measured with SMPS and OPC and fit using the sum of three lognormal distributions. The $dN/d(\log D_p)$ scale has units of cm^{-3} .
- Figure 4.** Effect of ozone level and air-exchange rate on steady-state fine particle ($\text{PM}_{1.1}$) mass concentration and yield. The lines are drawn to guide the eye. Solid symbols correspond to an air-exchange rate (AER) of 3 h^{-1} and hollow symbols indicate $\text{AER} = 1 \text{ h}^{-1}$.
- Figure 5.** Comparison of the measured change in particle number concentration with loss expected from ventilation alone for experiment POC-MH during stage 2. The time-dependent total SMPS particle number concentration is shown with a solid line, and lines are superimposed for apparent loss rate (dash) and loss expected from ventilation alone (dash-dot).
- Figure 6.** Modeled and measured particle mass concentrations vs. time for three experiments using the orange-oil degreaser. The sudden decrease in particle mass results from growth of the largest particles out of the size range of the SMPS.

Table 1. Experimental conditions. ^a

Exp. ID	Ozone level	Air-exchange rate	Supply ozone level (ppb)	Residual ozone level (ppb)	Air-exchange rate (h ⁻¹)	Supply terpene level (ppb)	Residual terpene level (ppb)
Orange-oil based degreaser (OOD)							
OOD-HH	High	High	137	21	3.0	643	518
OOD-HL	High	Low	136	7	1.0	738	588
OOD-MH	Moderate	High	61	11	3.1	586	528
Heated-oil air freshener (AFR)							
AFR-HH	High	High	126	18	3.0	623	492
AFR-HL	High	Low	127	4	1.0	859	658
AFR-MH	Moderate	High	63	7	3.0	596	506
Pine-oil based cleaner (POC)							
POC-VH	Very high	High	253	25	3.0	716	439
POC-HH1	High	High	131	13	3.0	771	560
POC-HH2	High	High	121	12	3.0	692	531
POC-HL	High	Low	130	8	1.0	735	599
POC-MH	Moderate	High	65	2	3.0	734	673
POC-LH	Low	High	29	0	3.1	566	540
POC-NOx	High	High	139	8	3.0	817	615
POC-Rev	High	High	120	-	3.0	-	-
POC-Seed	High	High	125	-	3.0	-	-
POC-Dry	Moderate	High	63	-	3.0	558	512

^a For experimental identification, the pre-dash letters denote the cleaning product and the post-dash letters denote some key aspect of experimental conditions, as follows: “HH” indicates high ozone level (~130 ppb at the inlet) and high air-exchange rate (AER = 3 h⁻¹); “HL” indicates high ozone level and low AER (1 h⁻¹); and “MH” indicates moderate ozone level (~60 ppb at the inlet) and high AER. Extra experiments were conducted with POC: “VH” denotes very high supply ozone (253 ppb at the inlet) and high AER; “HH1” and “HH2” are replicate experiments under HH conditions; “LH” indicates low ozone level (29 ppb at the inlet) and high AER. See text for a description of the remaining four POC experiments.

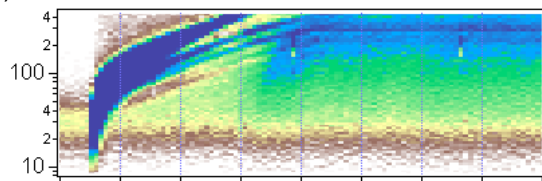
Table 2. Secondary aerosol size distribution parameters and formation rates.

Exp. ID	Measured						Modeled												
	Peak number conc. (10 ⁵ cm ⁻³)	Peak mass conc. (μg m ⁻³)		Steady-state number conc. (10 ⁵ cm ⁻³)	Steady-state mass conc. (μg m ⁻³)		Lognormal fit parameters									Steady-state mass conc. (μg m ⁻³)		R _{N,SS} (cm ³ s ⁻¹)	R _{M,SS} (μg m ⁻³ min ⁻¹)
							Mode 1			Mode 2			Mode 3						
		PM _{0.4}	PM _{1.1}		N (cm ⁻³)	GMD (nm)	GSD	N (cm ⁻³)	GMD (nm)	GSD	N (cm ⁻³)	GMD (nm)	GSD	PM _{0.4}	PM _{1.1}				
OOD-HH	1.4	229	-	0.18	105	-	2300	48	1.6	7000	135	1.6	8700	314	1.4	126	259	15	13
OOD-HL*	0.8	192	-	0.13	71	-	1500	40	1.5	5000	105	1.5	6000	310	1.4	87	162	4	3
OOD-HL*				0.10	87	-	1500	70	1.6	2500	140	1.5	7500	330	1.4	99	243	3	4
OOD-MH	0.6	75	-	0.13	53	-	2200	45	1.6	5700	123	1.6	4700	265	1.4	63	92	11	5
AFR-HH	0.4	51	93	0.08	33	80	2200	45	1.6	3200	125	1.5	2700	280	1.35	39	54	7	3
AFR-HL	0.9	90	146	0.05	37	110	1300	55	1.7	1700	140	1.5	3300	300	1.4	44	83	1	1
AFR-MH	0.2	21	35	0.07	15	29	1600	35	1.55	2900	100	1.6	1800	222	1.4	18	21	6	1
POC-VH	3.6	306	435	0.16	107	215	3000	45	1.6	5000	130	1.6	7500	315	1.4	104	220	13	11
POC-HH1	2.3	174	273	0.12	67	115	2500	48	1.6	4500	140	1.6	4700	316	1.4	72	147	10	7
POC-HH2	2.4	162	-	0.11	66	-	3000	60	1.7	3080	138	1.5	6000	295	1.4	79	144	9	7
POC-HL*	1.6	102	-	0.14	58	-	1500	80	1.7	5000	140	1.4	6000	243	1.3	69	75	4	1
POC-HL*				0.10	75	-	3000	70	1.7	1400	211	1.35	4900	305	1.3	78	111	3	2
POC-MH	1.2	57	-	0.24	50	-	1300	25	1.5	13000	75	1.7	7000	223	1.4	64	76	20	4
POC-LH	0.5	12	35	0.1	11	19	4000	42	1.65	3800	110	1.56	1400	182	1.45	14	15	8	1
POC-NOx	1.8	146	215	0.14	64	185	2900	37	1.55	5500	115	1.6	5000	300	1.4	70	130	12	7
POC-Rev	1.9	149	220	0.12	67	175	2000	40	1.6	4500	120	1.6	6000	295	1.4	80	145	10	7
POC-Seed	3.2	141	225	0.28	100	212	9000	55	1.75	7000	105	1.5	15000	225	1.5	130	200	23	10
POC-Drv	0.9	62	-	0.06	31	-	1300	60	1.8	2000	130	1.5	3000	275	1.4	38	60	5	3

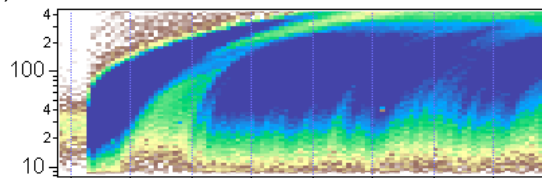
* Indicates experiments where the size distribution varied significantly during stage 4 and two sets of distribution parameters are given to show maximum and minimum values.

Figure 1

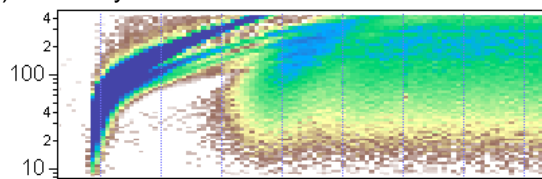
(a) POC-HH



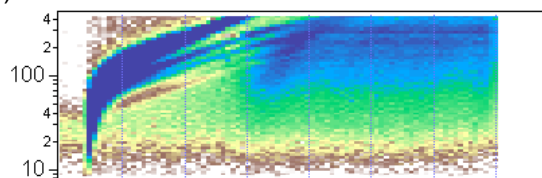
(b) POC-MH



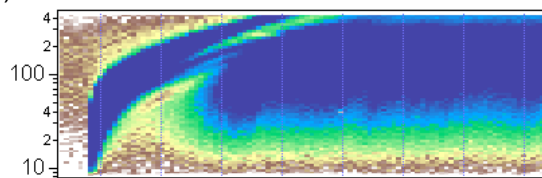
(c) POC-Dry



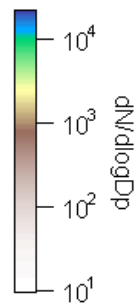
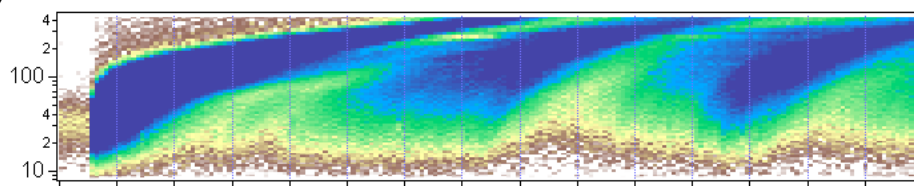
(d) POC-Rev



(e) POC-Seed



(f) POC-HL



Time (half-hour increments are marked)

Figure 2

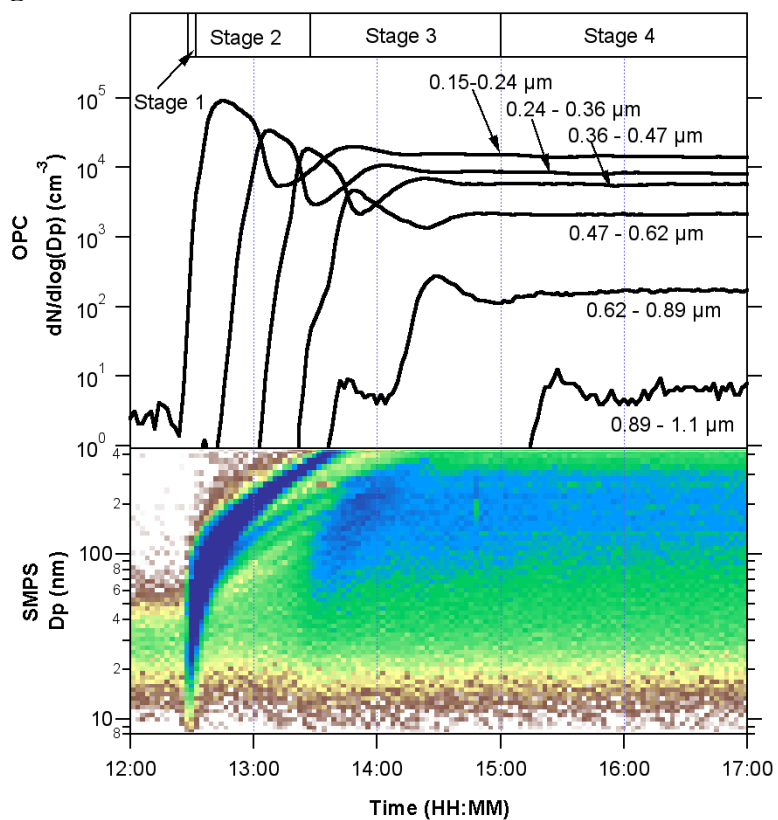


Figure 3

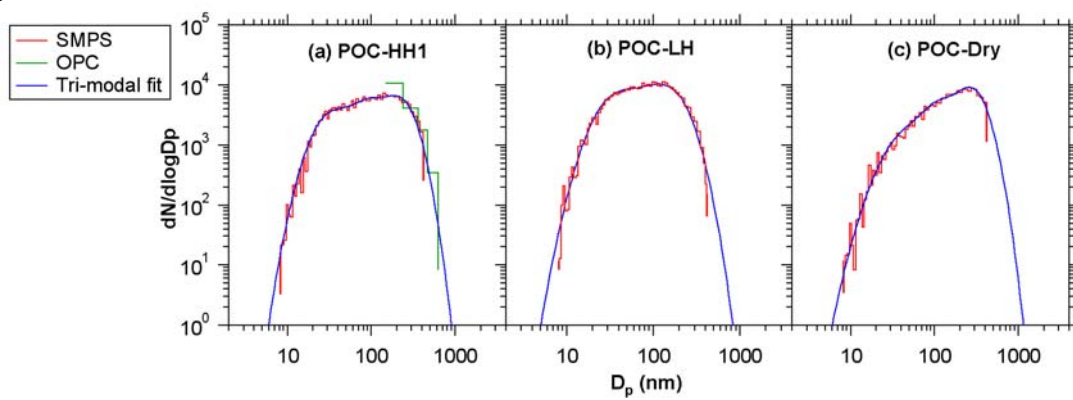


Figure 4

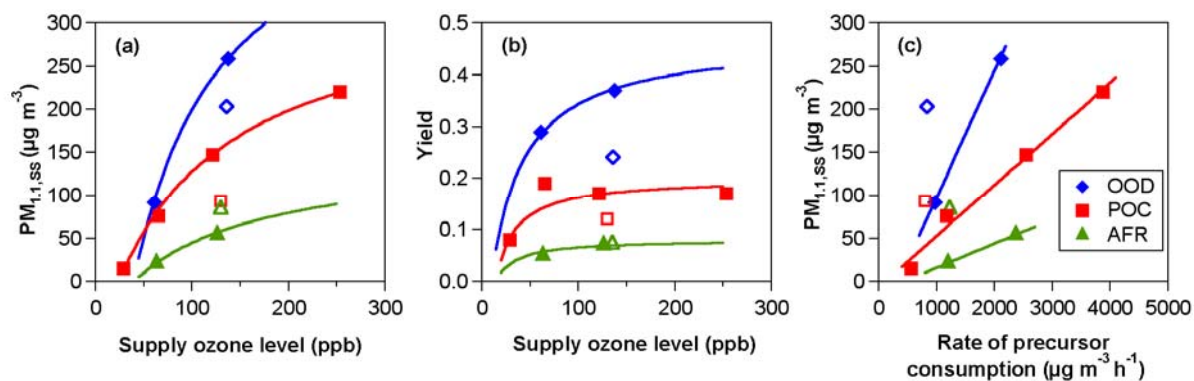


Figure 5

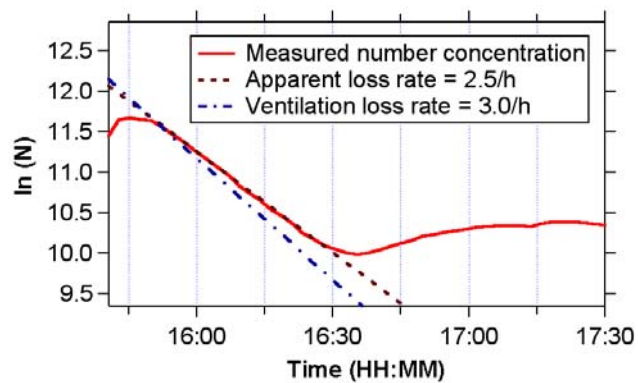


Figure 6

